

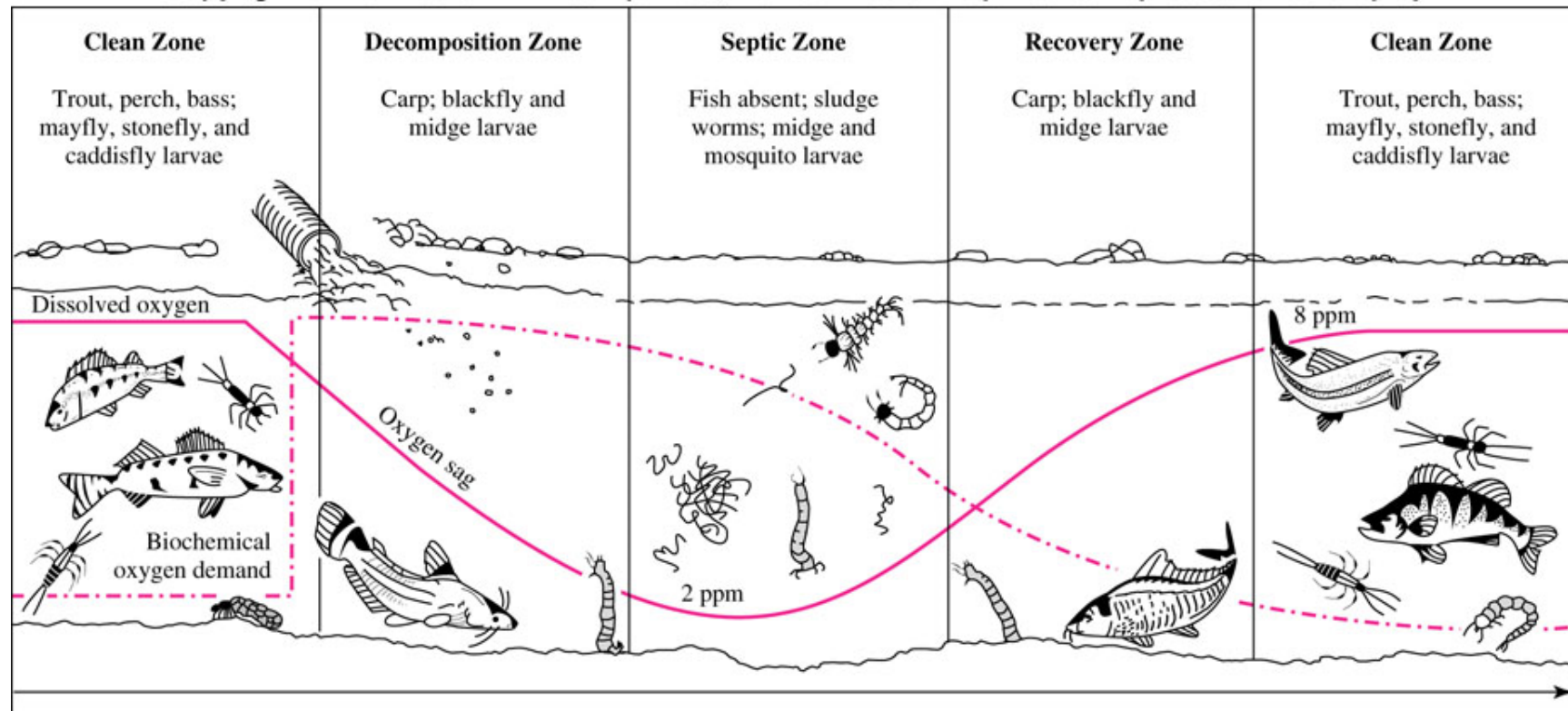
Water quality II

Water quality

- DO dynamics in river
 - DO sag curve
 - Modeling DO in a river
 - Solution: Streeter-Phelps equation
- Groundwater quality
 - Contamination issues
 - Contaminant transport mechanisms

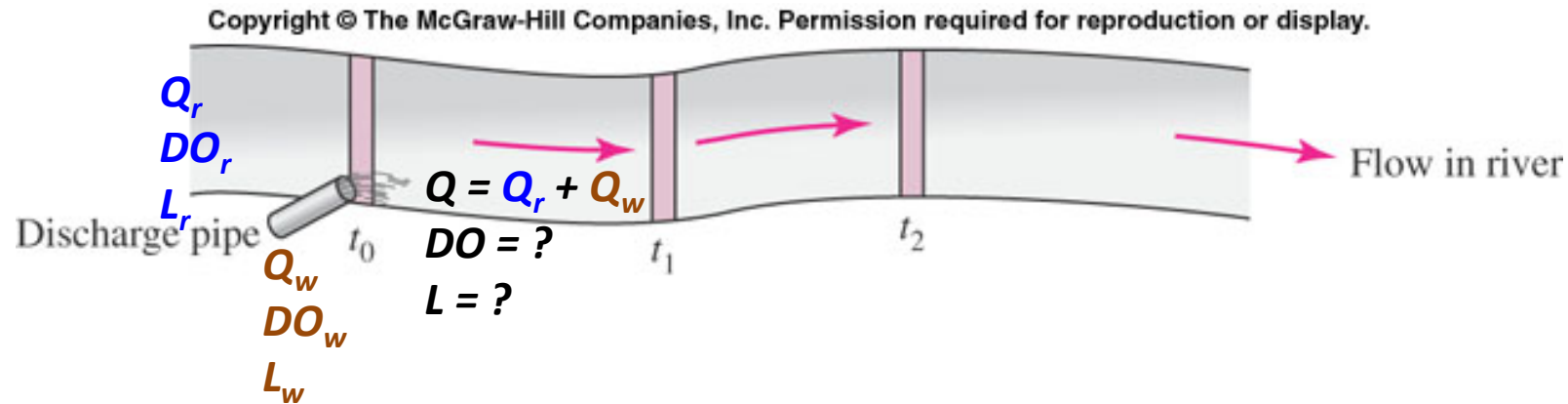
Water quality in rivers: DO sag curve

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- Factors causing DO depletion: BOD in water (upstream + waste)
- Factors causing DO increase: reaeration from the atmosphere (+ photosynthesis – neglected)

Modeling the DO along a river

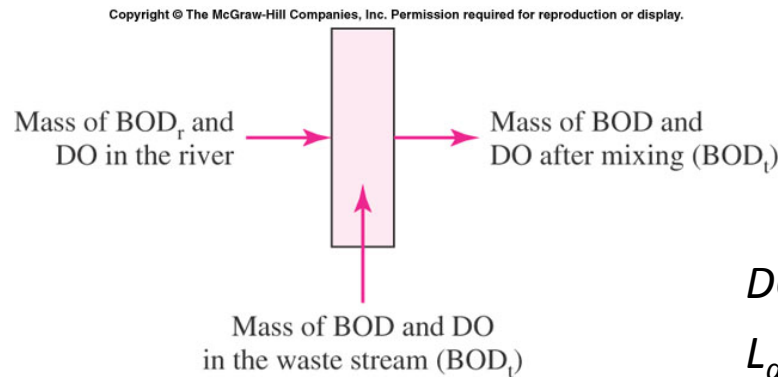


We will model the DO of a river receiving waste at time t_0 . The river will be modeled as a PFR.

* The solution for this problem is known as “Streeter-Phelps equation”, a well-known equation derived by Streeter and Phelps in 1925.

Modeling the DO along a river

The DO and ultimate BOD at t_0 are calculated by a mass balance approach:



$$(Q_w + Q_r)DO_a = Q_wDO_w + Q_rDO_r$$

$$(Q_w + Q_r)L_a = Q_wL_w + Q_rL_r$$

DO_a = DO concentration right after mixing (mg/L)

L_a = ultimate BOD right after mixing (mg/L)



$$DO_a = \frac{Q_wDO_w + Q_rDO_r}{Q_w + Q_r}$$

$$L_a = \frac{Q_wL_w + Q_rL_r}{Q_w + Q_r}$$

Modeling the DO along a river

The temperature after mixing is calculated in the same way:

$$T_a = \frac{Q_w T_w + Q_r T_r}{Q_w + Q_r} \quad T_a = \text{temperature after mixing (}^\circ\text{C or K)}$$

Oxygen deficit

- Oxygen deficit (D): the difference between the saturation DO value and the actual DO concentration

$$D = DO_s - DO$$

Therefore, the oxygen deficit right after mixing is calculated as:

$$D_a = DO_s - \frac{Q_w DO_w + Q_r DO_r}{Q_w + Q_r}$$

D_a = oxygen deficit right after mixing
(mg/L)

Modeling DO along a river

- Rate of reaeration

- Should depend on the stream velocity and depth

- The reaeration coefficient, k_r [day⁻¹]

$$k_r = \frac{3.9u^{1/2}}{h^{3/2}} \quad \begin{array}{l} u = \text{average stream velocity (m/s)} \\ h = \text{average stream depth (m)} \end{array}$$

- Rate of reaeration should also depend on oxygen deficit

$$\text{Rate of reaeration} = \left. \frac{d(DO)}{dt} \right|_{\text{reaeration}} = - \left. \frac{dD}{dt} \right|_{\text{reaeration}} = k_r D$$

- Effect of temperature on k_r : faster mass transfer at higher temp.

$$k_{r,T} = k_{r,20} \theta^{T-20} \quad \begin{array}{l} k_{r,T} = \text{reaeration coeff. at temperature } T \text{ (day}^{-1}\text{)} \\ k_{r,20} = \text{reaeration coeff. at } 20^\circ\text{C, obtained from} \\ k_{r,20} = 3.9u^{1/2}/h^{3/2} \text{ (day}^{-1}\text{)} \\ \theta = \text{temperature coefficient (use 1.024)} \end{array}$$

Modeling DO along a river

- Rate of deoxygenation
 - Rate of oxygen consumption by microorganisms
 - Assume that the first-order deoxygenation rate constant is equal to the BOD rate constant, k
 - The assumption is valid for deep, slow-moving streams
 - The rate of deoxygenation

$$\begin{aligned} \text{Rate of deoxygenation} &= -\left. \frac{d(DO)}{dt} \right|_{\text{deoxygenation}} = \left. \frac{dD}{dt} \right|_{\text{deoxygenation}} \\ &= k_d L \end{aligned}$$

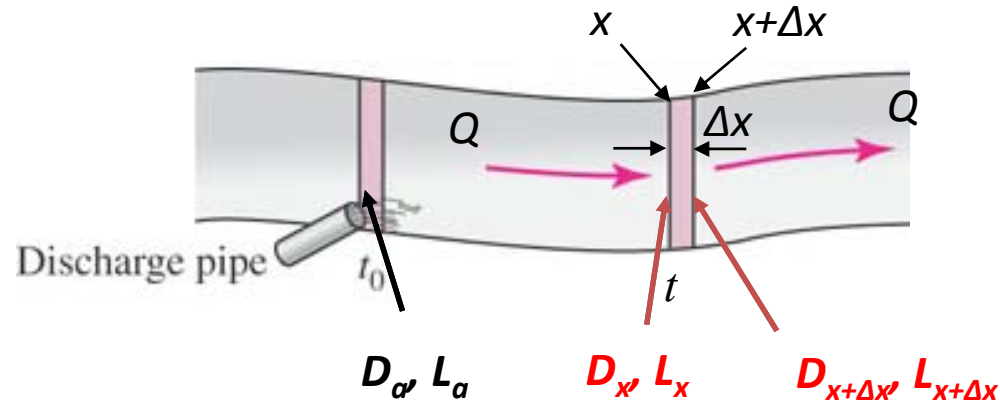
k_d = first-order deoxygenation rate constant [T⁻¹]

- Effect of temperature on k_d : use the equation for k !

$$k_T = k_{20} \theta^{T-20}$$

$\theta = 1.135$ for 4-20°C and
1.056 for 20-30°C

Modeling DO along a river



Steady-state D ($=DO_s - DO$) balance for a thin plate at time t :

$$0 = QD_x - QD_{x+\Delta x} + k_d L_x \cdot \Delta V - k_r D_x \cdot \Delta V \quad \Delta V = \text{volume of the CV} = A \cdot \Delta x$$

(A = cross-sectional area)

With rearrangements and $\Delta x \rightarrow 0$, we obtain:

$$\frac{dD}{dt} = k_d L - k_r D$$

Modeling DO along a river

Governing equation: $\frac{dD}{dt} = k_d L - k_r D$

+ Initial conditions:

at $t=0$, $D=D_a$ and $L=L_a$

Solution:

$$D_t = \frac{k_d L_a}{k_r - k_d} (e^{-k_d t} - e^{-k_r t}) + D_a (e^{-k_r t})$$

D_t = oxygen deficit in a river after flowing downstream from the mixing point for time t

(Note $DO_t = DO_s - D_t$)

Critical point

- Critical point: the point where the DO is the lowest on the DO sag curve

$$t_c = \frac{1}{k_r - k_d} \ln \left[\frac{k_r}{k_d} \left(1 - D_a \frac{k_r - k_d}{k_d L_a} \right) \right]$$

t_c = the time to the critical point [T]

- The critical deficit, D_c

$$D_c = \frac{k_d L_a}{k_r - k_d} (e^{-k_d t_c} - e^{-k_r t_c}) + D_a (e^{-k_r t_c})$$

Modeling DO along a river

Q: A city disposes of $1.05 \text{ m}^3/\text{s}$ of treated sewage having ultimate BOD of 28.0 mg/L and DO of 1.8 mg/L into a river. At the upstream from the outfall, the river flowrate is $7.08 \text{ m}^3/\text{s}$, and the ultimate BOD and DO of the river are 3.6 and 7.6 mg/L , respectively. At the river temperature, the saturation value of DO is 8.5 mg/L , the deoxygenation coefficient, k_d , 0.61 day^{-1} , and the reaeration coefficient, k_r , 0.76 day^{-1} . The velocity of the river downstream from the outfall is 0.37 m/s .

- 1) Calculate the ultimate BOD and DO just downstream from the outfall. Assume complete mixing.
- 2) Calculate the DO 16 km downstream from the outfall.
- 3) Calculate the critical time, distance, and the minimum DO.

Groundwater quality

- Contamination of groundwater (aquifer) can result from:
 - Discharge from improperly operated or located septic systems
 - Leaking underground storage tanks (USTs)
 - Improper disposal of hazardous and other chemical wastes
 - Spills from pipelines or transportation accidents
 - Recharge of groundwater with contaminated surface water
 - Leaking dumps and landfills
 - Leaking retention ponds or lagoons



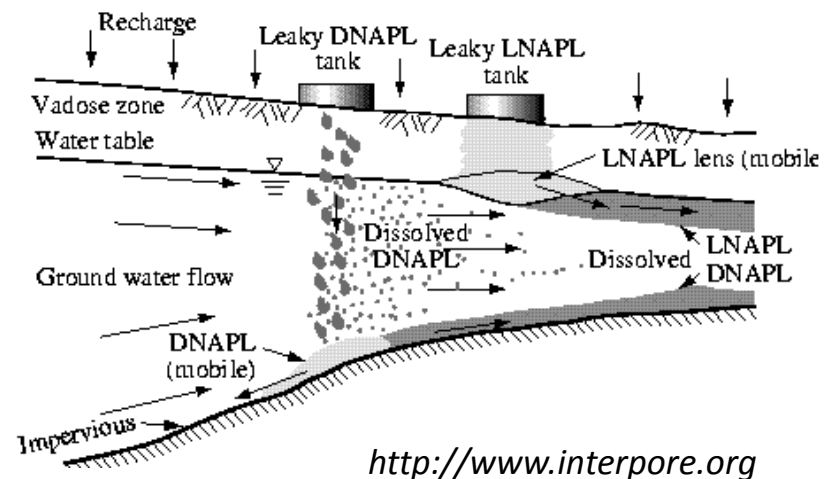
<http://www.septicrepairny.com>



<http://www.apexenvirotech.com>

Groundwater quality

- Nonaqueous phase liquid (**NAPL**) in aquifer
 - Many chemicals are only sparingly soluble in water
 - They may migrate in aquifer as a separate nonaqueous phase
 - **Light NAPL (LNAPL)**
 - lighter than water, float on the water table
 - example: gasoline (includes BTEX)
 - **Dense NAPL (DNAPL)**
 - Denser than water, sink in the aquifer until reaching an impermeable layer
 - Example: TCE, PCE



Groundwater quality

- Transport of dissolved contaminants
 - **Advection**: transport of dissolved contaminants by average movement of groundwater (seepage velocity)
 - **Dispersion**: spreading of contaminants by i) deviation of groundwater velocity from average and ii) molecular diffusion
 - Many contaminants move slower than the groundwater seepage velocity because of: **sorption** (adsorption + absorption)
 - Retardation coefficient: the extent to which chemicals are retarded in water

$$R = \frac{v'_{water}}{v'_{cont}}$$

R = retardation coefficient

v'_{water} = seepage velocity of groundwater

v'_{cont} = linear velocity of contaminant

Retardation coefficient

For neutral hydrophobic organic contaminants, the retardation coefficient, R , can be obtained by

$$R = 1 + \left(\frac{\rho_b}{\eta} \right) K_d$$

ρ_b = bulk density of soil (g/cm³)
 η = porosity of soil
 K_d = sorption coefficient of the contaminant between soil and water (cm³/g)
= (conc. in soil at equilibrium) / (conc. in water at equilibrium)

As hydrophobic organic contaminants mainly sorb to organic matter in soil, the K_d can be written as

$$K_d = K_{oc} \cdot f_{oc}$$

K_{oc} = sorption coefficient to the organic carbon fraction of soil (g/cm³)
= (conc. in organic carbon at equilibrium) / (conc. in water at equilibrium)
 f_{oc} = fraction of organic carbon in soil

Thus,

$$R = 1 + \left(\frac{\rho_b}{\eta} \right) K_{oc} \cdot f_{oc}$$

Transport of contaminants in groundwater

Q: A plume of benzene is migrating in groundwater flowing at a seepage velocity of 4.7×10^{-6} m/s. Using the following properties of the aquifer material and benzene, calculate the time for the center of the benzene plume to move 10 m in the direction of groundwater flow.

Aquifer material properties

Bulk density: 1.5 g/cm³

Porosity: 0.4

Fraction of organic carbon: 0.02

Benzene property

Sorption coefficient to the organic carbon fraction: 27.0 cm³/g

Reading assignment

Textbook Ch 9 p. 403-420, 435-439